



The Dilemma of Maintaining Intact Forest Through Certification

Fritz Kleinschroth^{1*}, Tim Rayden² and Jaboury Ghazoul^{1,3,4}

¹ Ecosystem Management Group, Department of Environmental Systems Science, ETH Zurich, Zurich, Switzerland, ² Wildlife Conservation Society, Bronx, NY, United States, ³ Prince Bernhard Chair for International Nature Conservation, Ecology and Biodiversity Group, Utrecht University, Utrecht, Netherlands, ⁴ Centre for Sustainable Forests and Landscapes, University of Edinburgh, Edinburgh, United Kingdom

Intact forests are natural and often extensive forests free from apparent anthropogenic degradation. Intact forests have important intrinsic and societal values, making their protection a high conservation priority. They are, however, vulnerable to being lost and degraded due to high opportunity costs and a lack of positive incentives to their preservation. Market-based mechanisms, such as voluntary certification, might provide a means to conserve intact forests while maintaining income through sustainable forest uses. Yet possibilities to ensure strict protection of large areas of intact forests through certification remain limited as long as premiums from certification are bound to the units of forest products that are sold. We explore challenges for incorporating intact forests into certification processes, and of maintaining intact forests within forest management units. To circumvent these challenges, it might be necessary to create a form of compensation payment scheme to overcome the foregone costs of intact forest preservation. Alternatively, certification systems might need to consider permitting some degree of regulated extraction in exchange for recognition and implementation of stringent forest preservation. This will require a re-evaluation of the way intactness is treated within current certification standards and the requirements for forestry within intact forests. Eventually, intact forest conservation and socially and economically viable forest management can only be reconciled on the landscape scale.

Keywords: land sharing land sparing, protected areas, REDD+, forest management, FSC, sustainable intensification, boreal forest, tropical forest

OPEN ACCESS

Edited by:

Yadvinder Malhi,
University of Oxford, United Kingdom

Reviewed by:

Marion Pfeifer,
Newcastle University, United Kingdom
Constance L. McDermott,
University of Oxford, United Kingdom

*Correspondence:

Fritz Kleinschroth
kifritz@ethz.ch

Specialty section:

This article was submitted to
Tropical Forests,
a section of the journal
Frontiers in Forests and Global
Change

Received: 14 December 2018

Accepted: 25 October 2019

Published: xx November 2019

Citation:

Kleinschroth F, Rayden T and
Ghazoul J (2019) The Dilemma of
Maintaining Intact Forest Through
Certification.
Front. For. Glob. Change 2:72.
doi: 10.3389/ffgc.2019.00072

INTRODUCTION

Global efforts for biodiversity conservation are not sufficient to be distributed equally around the world. In order to target those areas with the highest conservation value, two contrasting concepts have emerged, both prioritizing landscapes that are biodiverse, but one focussing more on hotspots the other more on coldspots of human activities. Hotspots are global centers of biological diversity and endemism that are threatened by human activity, especially from habitat loss (Brooks et al., 2002). Coldspots are extensive and largely intact and undisturbed natural regions where the threat of loss is less immediate, but where the problem of degradation is increasingly important. The maintained integrity of coldspots is important for their large carbon stores and the extensive habitats of many disturbance-sensitive species (Watson et al., 2018). According to recent research, areas that have been identified as global hotspots currently contain an average of only 15% of their natural, intact vegetation (Sloan et al., 2014). Coldspots, in contrast, include the last large intact

115 forests that remain free of human activities. Intact forests tend
 116 to be remote from populated areas and urban centers, and often
 117 occupy mostly inaccessible and agriculturally marginal regions in
 118 both tropical and boreal regions. One commonly used definition
 119 defines intact forest landscapes (IFL) as areas of at least 500
 120 km² that do not show any sign of remotely detectable human
 121 activity or habitat fragmentation (Potapov et al., 2008). Intactness
 122 is in itself a valued aspect of conservation quite apart from
 123 the biodiversity that such IFL might contain, and therefore
 124 preserving intactness is an additional and complementary
 125 component of conservation. The expansion of exploitative
 126 activities even into some of the most remote corners of the
 127 globe is stimulating efforts to maintain these extensive areas of
 128 permanent forest cover, especially in countries where pressure to
 129 harvest timber or convert forest to agricultural uses is high.

131 THE ECONOMIC IMPERATIVES OF USING 132 AND NOT USING INTACT FORESTS 133 134

135 Just as avoided deforestation is a cost-effective way for climate
 136 mitigation (Griscom et al., 2017), conserving intact forests
 137 has been described as a cost-effective way of delivering
 138 conservation benefits (Potapov et al., 2008). The underlying
 139 assumption is that maintaining an intact forest by avoiding
 140 human interventions of any kind has lower direct costs than
 141 maintaining, managing, or restoring smaller forested areas
 142 located in populated biodiversity hotspots. Large countries, such
 143 as Canada and Russia in boreal regions, or Brazil and the
 144 Democratic Republic of Congo in the tropics, could potentially
 145 maximize conservation outcomes for lower cost by preserving
 146 existing intact forests.

147 Yet, while on global scale the protection of intact forests can be
 148 a win to society, locally some people lose their assets. Many intact
 149 forests overlap with commercial logging interests (e.g., Courbin
 150 et al., 2014; Gaveau et al., 2014; Kleinschroth et al., 2017) and
 151 have been or will be exploited for timber under a business as
 152 usual scenario. The opportunity costs for avoided exploitation
 153 of resources within intact forests can be very high (Nasi et al.,
 154 2012). Areas of intact forests often have high commercial value
 155 for wood production, due to the age of forest stands, and the
 156 prevalence of large old trees. Forest companies have a strong
 157 financial interest to access the “primary forest premium,” and
 158 governments are attracted to the tax revenues generated from
 159 commercial logging. If governments do protect intact forest
 160 areas to the exclusion of extractive industries, some form of
 161 compensation payments (e.g., for ecosystem services) might be
 162 demanded by concession holders. Both REDD+ and mitigations
 163 for environmental impacts elsewhere could, theoretically, fund
 164 this. Yet, such compensation schemes are only viable if the funds
 165 are competitive with the expected extractive revenues (Butler
 166 et al., 2009). Additionally, in countries with limited statehood,
 167 characterized by weak institutional capacity in the periphery,
 168 the commitment to preserve forests might weaken over time, or
 169 might never materialize, as happened to Ngoyla-Mintom forest,
 170 one of the last intact forests outside national parks in Cameroon
 171 (Ongolo, 2015).

172 The long-term preservation of intact forests is also threatened
 173 by national development agendas. Nations typically seek to
 174 improve transport and power infrastructures in order to aid the
 175 extraction of natural and mineral resources, and reduce post-
 176 harvest losses in the food sector by increasing accessibility to rural
 177 lands. Logging is often a first step in this process, as it generates
 178 revenue and requires investment in initial infrastructure upon
 179 which subsequent development can be based.

181 FSC AS AN AGENT TO IMPLEMENT 182 INTACT FOREST CONSERVATION 183 184

185 Forest certification is a voluntary, market-based incentive
 186 mechanism to validate sustainable forest management for wood
 187 production in addition to legal compliance as a form of non-
 188 state governance. As such, it relies for its effectiveness on the
 189 marketing of forest products from responsibly managed forests.
 190 There is a need for market rewards to compensate owners
 191 for the cost of certification. The process of forest management
 192 certification implicitly follows a “land sharing” approach, based
 193 on the assumption that improved management across the whole
 194 management unit delivers overall benefit on social, environment
 195 and economic grounds.

196 FSC certification rules require a minimum of 10% of
 197 the management unit area to be set aside for conservation
 198 purposes (FSC Policy Standards Unit, 2010). In practice, this is
 199 complemented by areas designated as High Conservation Value
 200 (HCV) and un-operable areas, meaning that the percentage of
 201 protected forest within the management unit may be much
 202 higher. In 2014, FSC set itself the target to include IFL as an
 203 HCV criterion (FSC Policy Standards Unit, 2017), with far-
 204 reaching consequences for the implementation of certification
 205 in boreal and tropical forests (Kleinschroth et al., 2019). The
 206 opportunity costs introduced by the mandatory protection of
 207 IFL as part of certification depends on the individual location of
 208 a forest management concession and on the economic value of
 209 the IFL. The larger the overlap between concession and valuable
 210 IFL, the higher the opportunity costs (Karsenty and Ferron,
 211 2017). Intactness, as defined in the IFL concept, can only be
 212 maintained through strict protection. Yet, the price premium
 213 from certification is bound to the units of wood sold, not
 214 to the area protected (**Figure 1**). A company with concessions
 215 that include large overlaps with IFL areas will therefore be
 216 disadvantaged, unless it is compensated for the opportunity costs
 217 in a different way.

218 The influence of FSC over global IFL is small. In Russia, 1.6%
 219 of the 225 Mio ha of IFL area fell into certified concessions
 220 (Ptichnikov et al., 2017) and in the Congo Basin 1.2% of the 84
 221 Mio ha of IFL are found within certified concessions of Republic
 222 of Congo, Gabon, and Cameroon (based on own calculations
 223 for 2016). Other major overlaps between IFL and FSC certified
 224 areas occur in Canada and Brazil, where we were unable to find
 225 complete spatial data of certified areas. Total certified area in
 226 the six main IFL countries has stagnated since 2014 (**Figure 2**).
 227 In Africa, for example, the area of FSC certified forest has
 228 declined by 9% (from 7,421,322 to 6,784,372 ha) from February

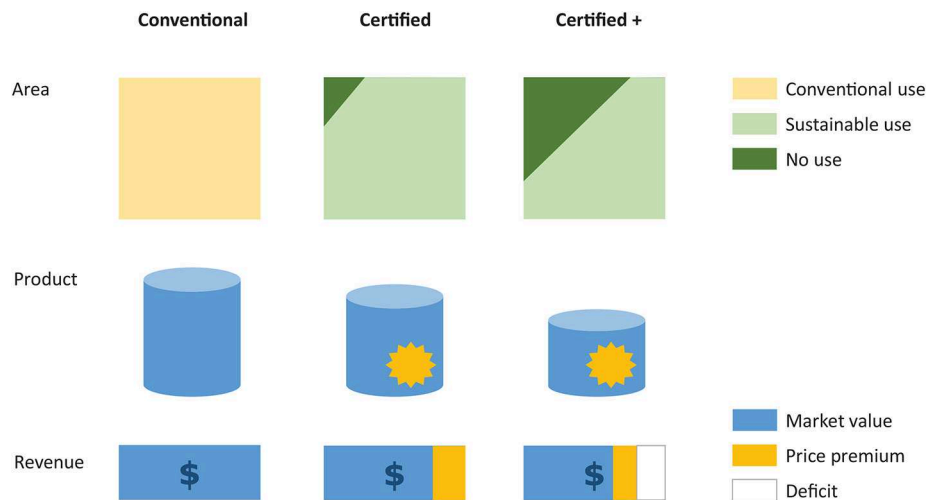


FIGURE 1 | Conceptual comparison between produced yields and income for the same area of forest under conventional and certified management. “Certified +” stands for certification that includes protection of IFL.

2016 to 2019 (<https://fsc.org/en/page/facts-figures>). This reflects, at least in part, the current atmosphere of uncertainty in the forestry sector regarding FSC certification to which the new IFL policy is contributing (Rotherham, 2016). Obtaining forest certification is a long process, and considerations as to whether to maintain a certificate might last longer than a few years, and such considerations might therefore not yet be reflected in currently reported certified areas. More remarkable is the strong increase of the area under double certification by FSC and the competing scheme Programme for the Endorsement of Forest Certification (PEFC). Data published jointly by FSC and PEFC shows a strong increase of the area under double certification in the three main IFL-countries Brazil, Canada, and Russia, as well as in all other countries from 2017 to 2018 (Figure 3). Around 43% of all FSC certified forest in Canada is now also certified by PEFC, with equivalent values being 27% for Russia and 51% for Brazil. This can be interpreted as a signal that the industry is seeking a backstop solution through an alternative certification scheme in the event that FSC is no longer tenable for them.

INCORPORATING IFL WITHIN THE FSC VOLUNTARY FRAMEWORK

The voluntary nature of certification means that the standards can only be as demanding as the marginal value of the certified label to the certified company. If standards become too demanding, certification will be a net cost, rather than a benefit to timber companies, resulting in “flight” from FSC. In order to prevent this, and to remain a viable influence in the timber trade, FSC could take either an “Exclusion Strategy,” abrogating responsibility by excluding intact forests from certified areas, or a “Reduced-impact Strategy,” allowing timber production in intact forests while attempting to reduce the impact of this activity with additional requirements. For other strategies

to become viable, certification would need to move further to a landscape scale, as we propose in the last section of this article.

The Exclusion Strategy excludes forest concessions that overlap with intact forests from certification, by not allowing any new certificates in IFL areas or by revoking existing certificates from IFL areas. The FSC has been criticized for certifying logging inside intact forests (Greenpeace, 2017). Removing IFLs from the certified area protects the reputation of the FSC brand at a superficial level, but fails to address the drivers of intact forest degradation. From a conservation perspective, the Exclusion Strategy is only useful if areas excluded from FSC certified forestry operations are also excluded from any other uses and become protected by governments. Yet, protected areas around the world experience strong human pressures (Jones et al., 2018; Schulze et al., 2018) and the effectiveness of strict protected area management is limited (Oldekop et al., 2016). Furthermore, while the wider implementation of REDD+ payments remains deadlocked (see e.g., Nantongo and Vatn, 2019), governments have few if any sources of compensation for the creation of additional protected areas. The likely outcome is that many IFLs would be exploited by companies using conventional (non-certified) harvesting methods, or companies using other certificates that lack IFL considerations (Karsenty and Ferron, 2017).

Alternatively, FSC could follow the Reduced-Impact Strategy, in the expectation that impacts on intact forests would be much less under light and highly regulated extractive management than alternative exploitation scenarios. This approach would allow timber extraction from an agreed portion of IFL areas within certified concessions, on the basis of tighter requirements on timber harvesting practice, post logging controls and increased permanent conservation set asides in critically important areas. This would allow FSC to govern actions

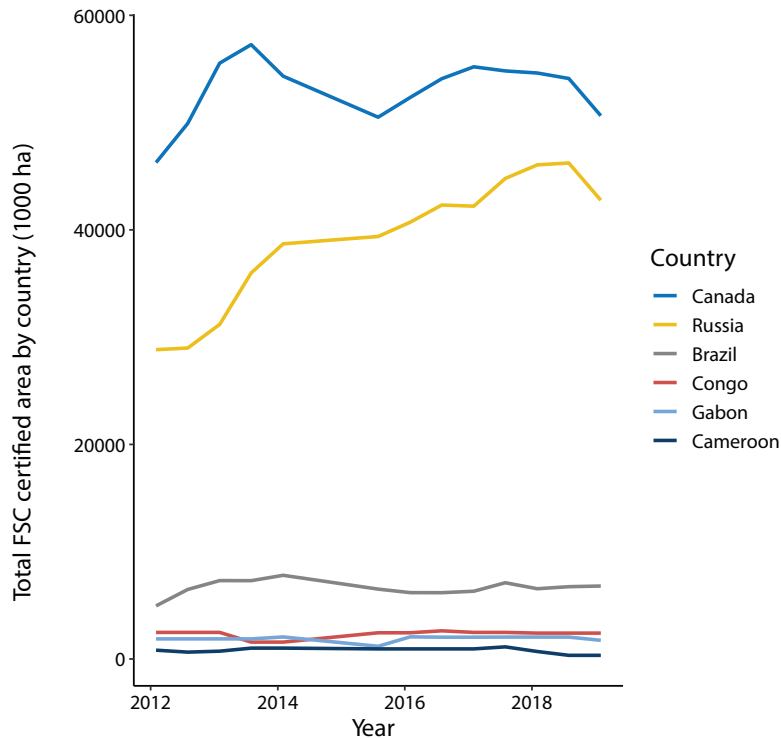


FIGURE 2 | Total FSC certified area between 2012 and 2019 in six countries with large IFL (source: <https://fsc.org/en/page/facts-figures>).

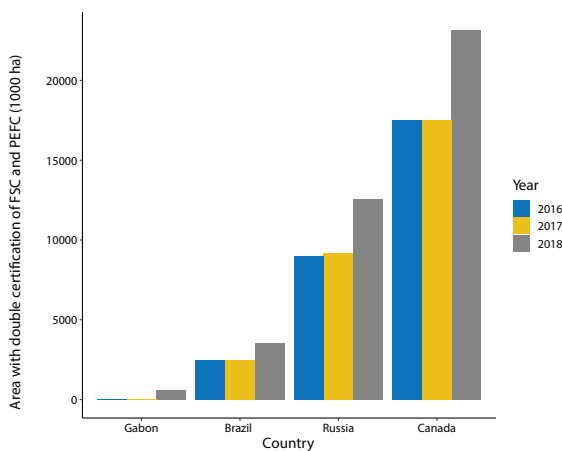


FIGURE 3 | Area (ha) with double certification by FSC and PEFC for countries with large IFL (sources: <https://fsc.org/en/page/facts-figures>, <https://www.atibt.org/en/press-release-of-our-partner-pacf-gabon-the-forest-certification-pacf-gabon-continues-its-development-and-commitment-to-the-gabonese-forests-by-rose-ondo-president-of-pacf-gabon/>).

forest stand due to tree harvesting (Martin et al., 2015), and provision of access to other land uses due to road building (Kleinschroth and Healey, 2017). Both processes can have severe impacts on plant and animal communities. Forest recovery strategies should, therefore, be an integral part of any forest management considerations. Forest recovery strongly depends on logging intensity (Kleinschroth et al., 2013). Common logging cycles of 30 years are considered too short to sustain yields of commercial species (Karsenty and Gourlet-Fleury, 2006), resulting in the strong contrast in standing value between intact and logged forests. At the same time carbon stocks in managed Amazonian forests have been shown to recover within <21 years at logging intensities below 30 m³ ha⁻¹ (Rutishauser et al., 2015). For disturbance sensitive animal species such as the woodland caribou (*Rangifer tarandus caribou*) in Canada, habitat recovery after clearcutting forestry operations takes at least 50 years (Environment Canada, 2012). In contrast, populations of chimpanzees (*Pan troglodytes troglodytes*) and gorillas (*Gorilla gorilla gorilla*) in tropical managed forests returned to baseline within <10 years after logging (Morgan et al., 2017). Especially in Central Africa, well-managed forests make an important contribution to species conservation (Clark et al., 2009; Stokes et al., 2010; Poulsen et al., 2011; Maisels et al., 2013). Forest and species recovery after logging are highly variable depending on geographical contexts. This highlights the importance of regional assessments of forest intactness to be used in forest management standards implemented on the ground.

The current identification of IFL is based on remote sensing and the most visible traces of industrial logging are the roads that are constructed for access. Definitions of intactness that could work within the FSC system could take into account ecological values on finer scales, and differentiate for the actual impact of different types of disturbance depending on the duration of time they occur, and on how quickly and effectively habitats can be restored afterwards. In such a case, the intact forest is maintained as an extensive forest unit that has not lost the main functions, carbon storage and the provision of habitat to disturbance sensitive species. The only forestry activities allowed would be constrained and regulated by strict adherence to FSC guidelines. Criteria for the definition of intact forests could include the extensiveness (e.g., more than 500 km²) of continuous cover forest with viable populations of monitored umbrella species such as the above mentioned woodland caribou in boreal regions and primates in the tropics. Further management would need to incorporate ecosystem service outcomes that are accommodated within a carefully managed and certified concession. Human activities would be limited to those permitted by the certifications standards, and any interventions (e.g., logging areas and roads) should no longer be discernible through remote sensing within 5 years of their implementation (Kleinschroth et al., 2015). The Reduced-impact Strategy presupposes that effective monitoring and verification of the efficacy of certification guidelines in maintaining biodiversity and ecosystem services and functions across intact forest areas.

THE CONSEQUENCES FOR FOREST MANAGEMENT IN NON-INTACT AREAS

If FSC requires companies to retain the oversight and management of IFLs, it is likely that timber production from these areas will have to be reduced in line with more stringent operational requirements, even to zero under current IFL standards. This may drive the intensification of timber extraction outside IFL areas to maintain current levels of timber output with implications for the implementation of certification standards and the marketing of certified wood. The potential for companies to do this while remaining within the standards expected of FSC certification remains uncertain, and will no doubt vary on a case by case basis, but the more general point is that we might expect pressure to increase on non-IFL forests within concession areas. Sustainable intensification is an approach to minimize the environmental footprint of productive systems by increasing outputs per area for multiple purposes (Rockström et al., 2017), but the extent to which this can be achieved in natural (i.e., not plantation) tropical or boreal forests has yet to be assessed in detail.

In clearcutting regimes of boreal forests, sustainable intensification would mean higher investments in silvicultural interventions before and after harvesting, requiring investment from forestry companies in technology, recruitment and training of skilled employees (Naumov et al., 2016). For tropical forests with selective logging regimes, intensification could be achieved through higher extraction regimes in previously disturbed

forest to increase light availability for faster regeneration of light-demanding timber species (Fredericksen and Putz, 2003). Other improvements include more careful mapping and planning processes, and using remote sensing and precision forestry before any operations take place. Increased efficiency in timber processing and transformation to reduce waste, and the marketing of a broader range of species, offer additional options for sustainable intensification (Karsenty et al., 2008; Horne, 2013). Current forestry regimes in remote regions with low tenure security may not, however, favor intensification on account of the costs relative to the returns when compared to conventional logging systems (Mathey et al., 2008).

URGENT NEED FOR LANDSCAPE SCALE SOLUTIONS

Land use changes in increasingly remote regions push back the forest frontier through degradation and forest clearance, and increased vulnerability to fire and illegal encroachment (Ahrends et al., 2010). To preserve intact forests, expansion into the forest frontier needs to halt. Buffers of managed natural forests might have an important function in maintaining a stable frontier between intact forests and agricultural land (Gaveau et al., 2013), provided that these activities are genuinely sustainable, and managed in a way that does not facilitate “hidden” encroachment as has been observed in agroforests that expand into natural forests legally or otherwise.

Care should be taken to ensure that “Exclusion Strategies” do not lead to displacement of unsustainable forest uses to other areas or countries with weaker law enforcement (Lambin and Meyfroidt, 2011). Such leakage has been observed in the context of REDD+, where deforestation was avoided where it was been paid for, but this led to forest losses elsewhere (Fisher et al., 2011). FSC provides some leverage to protect more intact forest areas, while ensuring financial benefits flow to forests country governments. Yet, any effort of FSC to protect intact forests will be spatially limited to those areas where certified concessions overlap with intact forests. Intact forests are generally larger than certified forest areas, meaning that measures to afford permanent protection to intact forests still depend on the creation, financing and management of protected areas. If certified forest management is to play a major on-the-ground role in intact forest protection, forest management certification of intact areas should be spatially aligned with protected areas.

Moving certification from the concession to the landscape scale, allows thinking beyond the land sharing—land sparing paradigm. Land use allocations in forested landscapes that strike a balance between productivity and conservation have been proposed. In a case study in Borneo, setting aside two-thirds of the land as protected areas could potentially be compensated by the incomes from certified selective logging and wood fiber plantations on the remaining third of the land (Runting et al., 2019).

Yet the landscape approach demands a coordination process that operates above the concession scale. Coordinated planning

that encompass a range of degraded, productive and intact forests in order to direct optimal spatial configurations of forest uses and restoration is not currently possible through management unit based voluntary certification. Moreover, supply chains emanating from regions such as the Congo Basin are structured around specific timber commodities, and a business plan built around plantations, even if only a small proportion of the land, is not necessarily viable. The proposed differential land allocation solution requires action from a range of stakeholders, including governments, and new paradigms for land use planning and conservation finance.

The protection of intact forests is gaining momentum and support from society, but existing certified companies view the IFL issue as a challenge to their continued viability in important timber producing regions (Rotherham, 2016; Karsenty and Ferron, 2017). To protect more intact forest, we need to explore ways of overcoming the concerns of certified companies that are often the most progressive actors in IFL frontier areas. Since these companies agreed to be certified, we can assume that they have some degree of willingness to respect and enforce ecological considerations in response to the demands of their target markets. To bridge this challenge, we might need to compromise on the strict non-intervention IFL approach, while still retaining the core elements of its agenda, including the preservation of extensive forest areas, the biodiversity they contain, and the services they provide. Alternatively, incentives could be provided in the form of compensation payments for non-exploitation, and these can be within the context of landscape-level payments for ecosystem

services (Ghazoul et al., 2009). In other words, certification that includes the protection of IFL areas could make a company potentially eligible for REDD+ payments. Making a stronger link between the ecological necessities of intact forest protection and the economic possibilities of certification can eventually strengthen both, for the benefit of livelihoods in production and conservation forests.

DATA AVAILABILITY STATEMENT

Publicly available datasets were analyzed in this study. This data can be found here: <https://fsc.org/en/page/facts-figures>.

AUTHOR CONTRIBUTIONS

FK conceived of the paper, wrote the first draft, and coordinated the writing. TR provided important input to content and structure of the text. JG led the research and finalized the writing of the paper.

FUNDING

FK was funded by the DAFNE project of the European Union H2020 programme (Grant No. 690268).

ACKNOWLEDGMENTS

Thanks to Gabriela Wiederkehr for processing data for **Figure 2**. We are grateful to three reviewers for their constructive comments on an earlier version of the manuscript.

REFERENCES

- Ahrends, A., Burgess, N. D., Milledge, S. A., Bulling, M. T., Fisher, B., Smart, J. C., et al. (2010). Predictable waves of sequential forest degradation and biodiversity loss spreading from an African city. *Proc. Natl. Acad. Sci. U.S.A.* 107, 14556–14561. doi: 10.1073/pnas.0914471107
- Brooks, T. M., Mittermeier, R. A., Mittermeier, C. G., da Fonseca, G. A. B., Rylands, A. B., Konstant, W. R., et al. (2002). Habitat loss and extinction in the hotspots of biodiversity. *Conserv. Biol.* 16, 909–923. doi: 10.1046/j.1523-1739.2002.00530.x
- Butler, R. A., Koh, L. P., and Ghazoul, J. (2009). REDD in the red: palm oil could undermine carbon payment schemes. *Conserv. Lett.* 2, 67–73. doi: 10.1111/j.1755-263X.2009.00047.x
- Clark, C. J., Poulsen, J. R., Malonga, R., and Elkan, P. W. (2009). Logging concessions can extend the conservation estate for Central African tropical forests. *Conserv. Biol.* 23, 1281–1293. doi: 10.1111/j.1523-1739.2009.01243.x
- Courbin, N., Fortin, D., Dussault, C., and Courtois, R. (2014). Logging-induced changes in habitat network connectivity shape behavioral interactions in the wolf-caribou-moose system. *Ecol. Monogr.* 84, 265–285. doi: 10.1890/12-2118.1
- Environment Canada (2012). *Recovery Strategy for the Woodland Caribou (Rangifer tarandus caribou), Boreal Population, in Canada*. Species at Risk Act Recovery Strategy Series.
- Fisher, B., Lewis, S. L., Burgess, N. D., Willcock, S., Kerry Turner, R., Malimbwi, R. E., et al. (2011). Implementation and opportunity costs of reducing deforestation and forest degradation in Tanzania. *Nat. Clim. Chang.* 1, 161–164. doi: 10.1038/nclimate1119
- Fredericksen, T. S., and Putz, F. E. (2003). Silvicultural intensification for tropical forest conservation. *Biodivers. Conserv.* 12, 1445–1453. doi: 10.1023/A:1023673625940

- FSC Policy and Standards Unit (2017). *International Generic Indicators FSC-STD-60-004 V1-1 EN*. Bonn: Forest Stewardship Council. Available online at: <https://fsc.org/en/document-center/documents/738f4a5e-136e-4625-8564-31c745e50b20>
- FSC Policy and Standards Unit (2010). *FSC Forest Stewardship Standards: Structure, Content and Suggested Indicators*. Bonn: Forest Stewardship Council.
- Gaveau, D. L., Kshatriya, M., Sheil, D., Sloan, S., Molidena, E., Wijaya, A., et al. (2013). Reconciling forest conservation and logging in Indonesian Borneo. *PLoS ONE* 8:e69887. doi: 10.1371/journal.pone.0069887
- Gaveau, D. L., Sloan, S., Molidena, E., Yaen, H., Sheil, D., et al. (2014). Four decades of forest persistence, clearance and logging on Borneo. *PLoS ONE* 9:e101654. doi: 10.1371/journal.pone.0101654
- Ghazoul, J., Garcia, C., and Kushalappa, C. G. (2009). Landscape labelling: a concept for next-generation payment for ecosystem service schemes. *For. Ecol. Manage.* 258, 1889–1895. doi: 10.1016/j.foreco.2009.01.038
- Greenpeace (2017). *Cut From Congo*. Greenpeace International. Available online at: https://cdn.greenpeace.fr/site/uploads/2017/10/Greenpeace_IFL_report_final.pdf
- Griscom, B. W., Adams, J., Ellis, P. W., Houghton, R. A., Lomax, G., Miteva, D. A., et al. (2017). Natural climate solutions. *Proc. Natl. Acad. Sci. U.S.A.* 114, 11645–11650. doi: 10.1073/pnas.1710465114
- Horne, J. (2013). *A Guide to Lesser Known Tropical Timber Species*. Gland: WWF International. Available online at: <https://www.worldwildlife.org/publications/guide-to-lesser-known-tropical-timber-species>
- Jones, K. R., Venter, O., Fuller, R. A., Allan, J. R., Maxwell, S. L., Negret, P. J., et al. (2018). One-third of global protected land is under intense human pressure. *Science* 360, 788–791. doi: 10.1126/science.aap9565

- 685 Karsenty, A., Drigo, I., Piketty, M., and Singer, B. (2008). Regulating industrial
686 forest concessions in Central Africa and South America. *For. Ecol. Manage.* 256,
687 1498–1508. doi: 10.1016/j.foreco.2008.07.001
- 688 Karsenty, A., and Ferron, C. (2017). Recent evolutions of forest concessions
689 status and dynamics in Central Africa. *Int. For. Rev.* 19, 10–26.
690 doi: 10.1505/146554817822295957
- 691 Karsenty, A., and Gourlet-Fleury, S. (2006). Assessing sustainability of logging
692 practices in the Congo Basin's managed forests: the issue of commercial species
693 recovery. *Ecol. Soc.* 11:26. doi: 10.5751/ES-01668-110126
- 694 Kleinschroth, F., Garcia, C., and Ghazoul, J. (2019). Reconciling certification
695 and intact forest landscape conservation. *Ambio* 48, 153–159.
696 doi: 10.1007/s13280-018-1063-6
- 697 Kleinschroth, F., Gourlet-Fleury, S., Sist, P., Mortier, F., and Healey, J. R. (2015).
698 Legacy of logging roads in the Congo Basin: how persistent are the scars in
699 forest cover? *Ecosphere* 6:64. doi: 10.1890/ES14-00488.1
- 700 Kleinschroth, F., and Healey, J. R. (2017). Impacts of logging roads on tropical
701 forests. *Biotropica* 49, 620–635. doi: 10.1111/btp.12462
- 702 Kleinschroth, F., Healey, J. R., Gourlet-Fleury, S., Mortier, F., and Stoica, R. S.
703 (2017). Effects of logging on roadless space in intact forest landscapes of the
704 Congo Basin. *Conserv. Biol.* 31, 469–480. doi: 10.1111/cobi.12815
- 705 Kleinschroth, F., Schöning, C., Kung'u, J. B., Kowarik, I., and Cierjacks, A.
706 (2013). Regeneration of the East African timber tree *Ocotea usambarensis*
707 in relation to historical logging. *For. Ecol. Manage.* 291, 396–403.
708 doi: 10.1016/j.foreco.2012.11.021
- 709 Lambin, E. F., and Meyfroidt, P. (2011). Global land use change, economic
710 globalization, and the looming land scarcity. *Proc. Natl. Acad. Sci. U.S.A.* 108,
711 3465–3472. doi: 10.1073/pnas.1100480108
- 712 Maisels, F., Strindberg, S., Blake, S., Wittemyer, G., Hart, J., Williamson, E., et al.
713 (2013). Devastating decline of forest elephants in central Africa. *PLoS ONE*
714 8:e59469. doi: 10.1371/journal.pone.0059469
- 715 Martin, P. A., Newton, A. C., Pfeifer, M., Khoo, M., and Bullock, J.
716 M. (2015). Impacts of tropical selective logging on carbon storage and
717 tree species richness: a meta-analysis. *For. Ecol. Manage.* 356, 224–233.
718 doi: 10.1016/j.foreco.2015.07.010
- 719 Mathey, A.-H., Krcmar, E., Innes, J., and Vertinsky, I. (2008). Opportunities and
720 costs of intensification and clustering of forest management activities. *Can. J.*
721 *For. Res.* 38, 711–720. doi: 10.1139/X07-197
- 722 Morgan, D., Mundry, R., Sanz, C., Ayina, C. E., Strindberg, S., Lonsdorf, E.,
723 et al. (2017). African apes coexisting with logging: comparing chimpanzee
724 (*Pan troglodytes troglodytes*) and gorilla (*Gorilla gorilla gorilla*) resource
725 needs and responses to forestry activities. *Biol. Conserv.* 218, 277–286.
726 doi: 10.1016/j.biocon.2017.10.026
- 727 Nantongo, M., and Vatn, A. (2019). Estimating transaction costs
728 of REDD+. *Ecol. Econ.* 156, 1–11. doi: 10.1016/j.ecolecon.2018.
729 08.014
- 730 Nasi, R., Billand, A., and van Vliet, N. (2012). Managing for timber
731 and biodiversity in the Congo Basin. *For. Ecol. Manage.* 268, 103–111.
732 doi: 10.1016/j.foreco.2011.04.005
- 733 Naumov, V., Angelstam, P., and Elbakidze, M. (2016). Barriers
734 and bridges for intensified wood production in Russia:
735 insights from the environmental history of a regional logging
736 frontier. *For. Policy Econ.* 66, 1–10. doi: 10.1016/j.forpol.2016.
737 02.001
- 738 Oldekop, J. A., Holmes, G., Harris, W. E., and Evans, K. L. (2016). A global
739 assessment of the social and conservation outcomes of protected areas. *Conserv.*
740 *Biol.* 30, 133–141. doi: 10.1111/cobi.12568
- 741 Ongolo, S. (2015). On the banality of forest governance fragmentation: exploring
742 “gecko politics” as a bureaucratic behaviour in limited statehood. *For. Policy*
743 *Econ.* 53, 12–20. doi: 10.1016/j.forpol.2015.01.005
- 744 Potapov, P., Yaroshenko, A., Turubanova, S., Dubinin, M., Laestadius, L., Thies, C.,
745 et al. (2008). Mapping the world's intact forest landscapes by remote sensing.
746 *Ecol. Soc.* 13:16. doi: 10.5751/ES-02670-130251
- 747 Poulsen, J. R., Clark, C. J., and Bolker, B. M. (2011). Decoupling the effects of
748 logging and hunting on an afro-tropical animal community. *Ecol. Appl.* 21,
749 1819–36. doi: 10.1890/10-1083.1
- 750 Ptichnikov, A., Dunn, A., Yanitskaya, T., and Burnisheva, J. (2017). FSC-
751 certification as a main tool for protecting intact forest landscapes in Russia.
752 Moscow: FSC Russia.
- 753 Rockström, J., Williams, J., Daily, G., Noble, A., Matthews, N., Gordon,
754 L., et al. (2017). Sustainable intensification of agriculture for human
755 prosperity and global sustainability. *Ambio* 46, 4–17. doi: 10.1007/s13280-
756 016-0793-6
- 757 Rotherham, T. (2016). Forest certification: trends and turbulence. *Can. For.*
758 *Ind. Mag.* 20–23. Available online at: [https://www.woodbusiness.ca/forest-](https://www.woodbusiness.ca/forest-certification-in-canada-trends-and-turbulence-3039/)
759 [certification-in-canada-trends-and-turbulence-3039/](https://www.woodbusiness.ca/forest-certification-in-canada-trends-and-turbulence-3039/)
- 760 Runting, R. K., Ruslandi, R., Griscom, B. W., Struebig, M. J., Satar, M., Meijaard,
761 E., et al. (2019). Larger gains from improved management over sparing-sharing
762 for tropical forests. *Nat. Sustain.* 2, 53–61. doi: 10.1038/s41893-018-0203-0
- 763 Rutishauser, E., Baraloto, C., Blanc, L., Descroix, L., Sota, E. D., Kanashiro, M., et al.
764 (2015). Rapid tree carbon recovery in Amazonian logged forests. *Curr. Biol.* 25,
765 191–201. doi: 10.1016/j.cub.2015.09.059
- 766 Schulze, K., Knights, K., Coad, L., Geldmann, J., Leverington, F., Eassom, A., et al.
767 (2018). An assessment of threats to terrestrial protected areas. *Conserv. Lett.* 11,
768 1–10. doi: 10.1111/conl.12435
- 769 Sloan, S., Jenkins, C. N., Joppa, L. N., Gaveau, D. L., a, and Laurance, W. F. (2014).
770 Remaining natural vegetation in the global biodiversity hotspots. *Biol. Conserv.*
771 177, 12–24. doi: 10.1016/j.biocon.2014.05.027
- 772 Stokes, E. J., Strindberg, S., Bakabana, P. C., Elkan, P. W., Iyenguet, F. C.,
773 Madzoké, B., et al. (2010). Monitoring great ape and elephant abundance at
774 large spatial scales: measuring effectiveness of a conservation landscape. *PLoS*
775 *ONE* 5:e10294. doi: 10.1371/journal.pone.0010294
- 776 Watson, J. E. M., Evans, T., Venter, O., Williams, B., Tulloch, A., Stewart, C., et al.
777 (2018). The exceptional value of intact forest ecosystems. *Nat. Ecol. Evol.* 2,
778 599–610. doi: 10.1038/s41559-018-0490-x
- 779 **Conflict of Interest:** The authors declare that the research was conducted in the
780 absence of any commercial or financial relationships that could be construed as a
781 potential conflict of interest.
- 782 *Copyright* © 2019 Kleinschroth, Rayden and Ghazoul. This is an open-access article
783 distributed under the terms of the Creative Commons Attribution License (CC BY).
784 The use, distribution or reproduction in other forums is permitted, provided the
785 original author(s) and the copyright owner(s) are credited and that the original
786 publication in this journal is cited, in accordance with accepted academic practice.
787 No use, distribution or reproduction is permitted which does not comply with these
788 terms.